

GASEOUS EMISSIONS (NH₃, N₂O, CH₄ AND CO₂) FROM A BIOLOGICAL AEROBIC TREATMENT UNIT OF PIG SLURRY

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ABSTRACT

Biological aerobic treatment of pig slurry is one of the measures to limit land-application of manure. This treatment transforms the nitrogen compounds into N₂ by nitrification and denitrification. However, this treatment may be a source of other gaseous emissions (NH₃, CH₄, N₂O, CO₂) emitted during the treatment and the following management of the treatment by-products. Within this framework, NH₃, N₂O, CH₄ and CO₂ release from a biological aerobic treatment system of pig slurry was investigated under field conditions. Field quantification and integration of the NH₃, CH₄, N₂O, CO₂ emission rates within a model based on the by-product flow enabled us to estimate annual gas flows. The results showed that the emissions of CH₄, CO₂ and NH₃ were more important for the non-aerated effluents (raw slurry and separated solid fraction). The total gas emissions of NH₃, CH₄ and CO₂ were close to 480 kgN, 3400 kgC and 12700 kgC per year respectively. The majority of the emissions occurred during the storage of the separated solid fraction (59% of NH₃ and 65% of CO₂) and the raw slurry (29% of the NH₃ and 54% of CH₄). N₂O was only measured above the biological reactor with 2.6 kgN per year.

Keywords: ammonia, methane, nitrous oxide, aerobic treatment.

INTRODUCTION

According to the EU nitrate directive, the high pig production in Brittany (France) implies a slurry treatment for some farms that don't have enough land to respect the level of 170 kg.N.ha⁻¹.year⁻¹. Among the different treatments the aerobic treatment consists in transforming the slurry nitrogen into gas (N₂) by a nitrification/denitrification process. However, N₂O and NH₃ could be emitted during the slurry aeration while the management of the treatment by-products (aerated liquid manure, solid separated fraction, biological sludge and supernatant) can also produce gas emissions (NH₃, N₂O, CH₄, CO₂). This paper concerns field measurements of NH₃, N₂O, CH₄ and CO₂ emissions during aeration of pig slurry and storage of by-products of a biological treatment unit. The integration of the determined gaseous factors to a total model on product flow led to the evaluation of gas flows on the scale of the treatment unit.

MATERIALS AND METHODS

The study was conducted on a commercial pig operation. The aerobic treatment unit consisted of (i) a raw slurry storage (ii) a solid-liquid separation by centrifuge (iii) a biological aerobic reactor (iv) a aerated slurry decantation-storage (v) a supernatant lagoon and (vi) a biological sludge storage. The gaseous emissions from the aerobic reactor and from the storage of raw slurry, aerated slurry, liquid separated fraction and biological sludge were measured using the technique of the dynamic chamber (Peu, 1999). Gaseous emissions during storage of the solid separated fraction was measured by enclosing the heap in a large polyethylene structure drawn up by the use of a wind motor. The concentrations of CH₄, CO₂ and N₂O were measured either

by infrared detection (analyser URAS 14 and 10) or by FID/ECD gas chromatography (Varian star 3800). The ammonia concentration was determined by using trap bottles filled with sulphuric acid (0.5 N) followed by a distillation and a titration. The gaseous measures were carried out for 4 to 7 weeks, in autumn for the storage of raw slurry and biological sludge, in spring for the biological reactor and the storage of aerated slurry and in winter for the storage of solid separation fraction and the supernatant. In order to estimate annual gaseous flow we developed a mathematical model that integrates the daily variations of volume of pig slurry (Levasseur 1998a, 1998b and 2002) and biological treatment by-products (based on field experimental measures, Beline et al. 2001 and 2003), the livestock production management (number of sows and housing according to the advice of French livestock institutes). The gas emission flows are calculated on a step of daily time by applying the field experimental gaseous factors to the various product storages.

RESULTS AND DISCUSSION

The results show the impact of the treatment on the characteristics of the by-products compared to the raw slurry (table 1). The level of TAN and organic matter (TVS) of the solid separated fraction, about 4.8 gN.kg⁻¹ and 270 g.kg⁻¹ are higher than those of raw slurry while the other biological treatment by-products present less TAN or organic matter.

Table 1. Characteristics of pig slurry and biological treatment by-products (Mean values).

Characteristic	Type of manure					
	RS	SSF	LSF	AS	S	BS
Total Ammonium Nitrogen (TAN, gN.kg ⁻¹)	2.9	4.8	2.3	0.2	0.05	0.2
Total Nitrogen (gN.kg ⁻¹)	4.1	11	3.1	1.2	0.1	1.7
COD (gO ₂ .kg ⁻¹)	44	353	23	19	2	31
Total Solids (g.kg ⁻¹)	42	340	19	22	7	34
Total Volatile Solids (TVS, g.kg ⁻¹)	30	270	11	10	2	14
Total Suspended Solids (g.kg ⁻¹)	36	-	8	15	1	28

RS: raw slurry, SSF: solid separated fraction, LSF: liquid separated fraction entering the biological reactor, AS: aerated slurry, S: supernatant, LD: liquid decantation fraction, BS: biological sludge.

The gas emission factors of CH₄, CO₂, NH₃ and N₂O are presented in table 2.

Table 2. Gaseous emissions factors (Mean values, ND: no detection).

Gas	Storage/treatment					
	RS	SSF	ABT	AS	S	BS
NH ₃	6.7 gN.m ⁻² .d ⁻¹	28 gN.T ⁻¹ .d ⁻¹	ND	0.16 gN.m ⁻² .d ⁻¹	0.35 gN.m ⁻² .d ⁻¹	0.3 gN.m ⁻² .d ⁻¹
N ₂ O	ND	ND	0.014 gN.m ⁻³ .d ⁻¹	ND	ND	ND
CH ₄	49.8 gC.m ⁻³ .d ⁻¹	60 gC.T ⁻¹ .d ⁻¹	0.8 gC.m ⁻³ .d ⁻¹	7.6 gC.m ⁻³ .d ⁻¹	ND	5.8 gC.m ⁻³ .d ⁻¹
CO ₂	41.9 gC.m ⁻³ .d ⁻¹	820 gC.T ⁻¹ .d ⁻¹	11.4 gC.m ⁻³ .d ⁻¹	4.8 gC.m ⁻³ .d ⁻¹	ND	7.1 gC.m ⁻³ .d ⁻¹

RS: raw slurry, SSF: solid separated fraction, ABT: aerobic biological treatment, AS: aerated slurry, S: supernatant, BS: biological sludge.

The CH₄ emissions of pig slurry is higher than the maximum value of 35.8 gCH₄.m⁻³.d⁻¹. cited in previous studies (Martinez et al., 1995, 2003, Husted, 1993, 1994). This value may have been the result of the frequent pit agitation which amplifies the emissions after agitation (Figure 1). The high production of CH₄ and CO₂ of the solid separated fraction is probably due to the

high concentration of water and biodegradable organic matter, two favourable parameters for composting. These results point out the need to a better management of this product for reducing the CH_4 emissions. The CH_4 and CO_2 emissions of the biological sludge were similar to those of aerated slurry. The CH_4 and CO_2 emitted by the supernatant lagoon were null probably due to the unfavourable winter temperatures for transforming the low biodegradable organic matter (Safley and Westerman, 1989; Husted, 1993, 1994.)

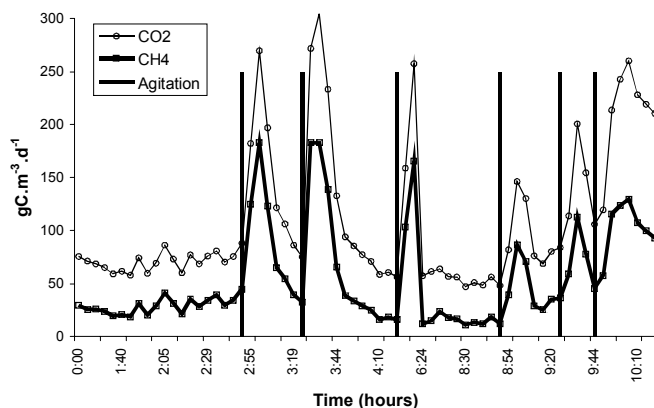


Figure 1. Influence of raw slurry (RS) agitation on gas emission (CH_4 and CO_2).

The NH_3 rates of raw slurry were relatively stable ($6.7 \text{ gN-NH}_3 \cdot \text{m}^{-2} \cdot \text{d}^{-1}$) despite the TAN variation. This NH_3 rate was also higher than those cited in the literature (De Bode, 1990; Sommer et al., 1993; Guingand, 2002). As for CH_4 , the pit agitation was probably the main reason of these considerable emissions. The high NH_3 emissions of the solid separated fraction ($28 \text{ gN-NH}_3 \cdot \text{T}^{-1} \cdot \text{d}^{-1}$) were related to the high TAN ($4.8 \text{ gN} \cdot \text{kg}^{-1}$) and the high temperature of the heap (65°C). The NH_3 emitted during the storage of the aerated slurry, biological sludge and the supernatant were low according to their low TAN content. N_2O emissions were only detected during the aeration of separated slurry with a mean value of $0.014 \text{ gN-N}_2\text{O} \cdot \text{m}^{-3} \cdot \text{d}^{-1}$. These results are consistent considering previous studies (Beline, 1998, Phillips and al., 1997, Osada and al., 1998) which described that N_2O production takes place mainly during the nitrification/denitrification process requiring aerobic and anaerobic conditions. The emission rates obtained in this study and the variance within the data set are mainly related to the difference of stored product characteristics. Indeed, the manure characteristics are one of the main parameters influencing the methane and ammonia emissions (De Bode, 1990; Husted, 1994; Sharpe, 2002). Under favourable conditions (temperature, pH...), the production of CH_4 is a function of the content of degradable organic matter while the NH_3 volatilisation is a function of the TAN. The higher these parameters are, the more important are the emissions of CH_4 and NH_3 . The annual emission rates were estimated by using the field-experimental data. We considered only the variation of the daily volume of the raw slurry and the treatment-by-products. The estimates presented in Table 3 concern a 200-sow production unit. These results indicate that the solid fraction storage is the main source of NH_3 with about 59% of the annual emissions, followed by the raw slurry storage (29%). The raw slurry are also the principal source of CH_4 with 54% of the annual emission. The gas emissions of the aerated by-product storages were less than for the raw slurry. N_2O was only measured above the biological reactor with 2.6 kgN per year.

CONCLUSIONS

This study confirms the influence of the composition and management of manure on gase-

ous emissions. The higher the ammonium and organic matter level are, the higher are the emissions of NH_3 , CH_4 and CO_2 . The annual estimates of NH_3 , CH_4 and CO_2 rate of this aerobic treatment unit were about 480 kgN, 3400 kgC and 12700 kgC respectively. The majority of the NH_3 and CH_4 emissions were from the solid separated fraction and the raw slurry. The main source of N_2O was the biological reactor. This study will be completed with a new field experiment to consider the seasonal effect on gas emissions.

Table 3. Annual gas estimations (ND: no detection).

Gas	Storage/treatment unit						Total
	RS	SSF	AT	AS	S	BS	
NH_3 kgN.y ⁻¹	139 (29%)	283 (59%)	ND	1.5 (0.3%)	44 (9%)	12 (3%)	480 (100%)
N_2O kgN.y ⁻¹	ND	ND	2.6 (100%)	ND	ND	ND	2.6 (100%)
CH_4 kgC.y ⁻¹	1854 (54%)	605 (18%)	143 (4%)	277 (8%)	ND	554 (16%)	3433 (100%)
CO_2 kgC.y ⁻¹	1560 (12%)	8273 (65%)	2047 (16%)	175 (1.5%)	ND	678 (5.5%)	12732 (100%)

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